

Cultivation of an obligate acidophilic ammonia oxidizer from a nitrifying acid soil

Laura E. Lehtovirta-Morley^a, Kilian Stoecker^b, Andreas Vilcinskas^b, James I. Prosser^{a,1}, and Graeme W. Nicol^a

^aInstitute of Biological and Environmental Sciences, University of Aberdeen, Aberdeen AB24 3UU, United Kingdom; and ^bFraunhofer Institute for Molecular Biology and Applied Ecology, 35394 Giessen, Germany

Edited by James M. Tiedje, Michigan State University, East Lansing, MI, and approved August 17, 2011 (received for review May 7, 2011)

Nitrification is a fundamental component of the global nitrogen cycle and leads to significant fertilizer loss and atmospheric and groundwater pollution. Nitrification rates in acidic soils (pH < 5.5), which comprise 30% of the world's soils, equal or exceed those of neutral soils. Paradoxically, autotrophic ammonia oxidizing bacteria and archaea, which perform the first stage in nitrification, demonstrate little or no growth in suspended liquid culture below pH 6.5, at which ammonia availability is reduced by ionization. Here we report the discovery and cultivation of a chemolithotrophic, obligately acidophilic thaumarchaeal ammonia oxidizer, "*Candidatus Nitrosotalea devanattera*," from an acidic agricultural soil. Phylogenetic analysis places the organism within a previously uncultivated thaumarchaeal lineage that has been observed in acidic soils. Growth of the organism is optimal in the pH range 4 to 5 and is restricted to the pH range 4 to 5.5, unlike all previously cultivated ammonia oxidizers. Growth of this organism and associated ammonia oxidation and autotrophy also occur during nitrification in soil at pH 4.5. The discovery of *Nitrosotalea devanattera* provides a previously unsuspected explanation for high rates of nitrification in acidic soils, and confirms the vital role that thaumarchaea play in terrestrial nitrogen cycling. Growth at extremely low ammonia concentration (0.18 nM) also challenges accepted views on ammonia uptake and metabolism and indicates novel mechanisms for ammonia oxidation at low pH.

acidophile | nitrification

Approximately 30% of the world's soils are acidic (pH < 5.5), including more than 50% of potential arable land (1). Although low pH reduces the rates of many soil ecosystem processes, a metastudy of gross nitrification in almost 300 soils showed weak negative correlation with soil pH and some of the highest rates in acid soils (2). This is paradoxical given the physiology of cultivated soil nitrifiers. Nitrification involves the oxidation of ammonia to nitrite and, subsequently, to nitrate. Ammonia oxidation in soil was traditionally considered to be dominated by autotrophic β -proteobacteria, but most cultivated bacterial ammonia oxidizers do not grow in suspended batch culture below pH 6.5 (3, 4). This is believed to be through reduced availability of NH_3 (pK_a for NH_3 : NH_4^+ , 9.25), the substrate for ammonia monooxygenase, which catalyzes conversion of ammonia to hydroxylamine (5, 6). In addition, in the absence of nitrite oxidation, ammonia oxidizers may be inhibited by nitric and nitrous acids, whose production from nitrite increases under acid conditions, and by generation of nitric oxide and nitrogen dioxide from nitrous acid, which is unstable and reactive at low pH.

Several mechanisms have been proposed to explain this paradox, including pH-neutral microenvironments, urease activity (7, 8), heterotrophic nitrification (9), biofilm and aggregate formation (10, 11), and close interactions between ammonifiers and ammonia oxidizers in soil aggregates (12). These mechanisms are difficult to demonstrate in soil and none fully explains high rates of nitrification in soils with pH less than 5.5. An additional explanation is the existence of acidophilic ammonia oxidizers, but all attempts to isolate such organisms have failed, although an enrichment was obtained at low pH (10) in which aggregate formation was required for ammonia oxidation. Attempts to cultivate acidophilic ammonia oxidizers preceded the discovery of ammo-

nia-oxidizing archaea, which fall within the thaumarchaeal lineage. These organisms appear to contribute significantly to ammonia oxidation in many soils in which thaumarchaeal *amoA* (ammonia monooxygenase subunit A) genes and transcripts outnumber equivalent bacterial genes (13–17). Their activity, relative to that of bacterial ammonia oxidizers, also appears to increase in soil pH transects with decreasing soil pH (13).

The aim of this study was to determine whether ammonia oxidation in acid soils resulted from the existence of acidophilic thaumarchaea by attempting their enrichment in low-pH, mineral salts medium containing ammonia, following inoculation with acid soil in which thaumarchaea are believed to drive nitrification (13, 14).

Results

Enrichment of an Obligate Acidophilic Thaumarchaeal Ammonia Oxidizer. To determine whether ammonia oxidation in acid soils was caused by acidophilic autotrophic ammonia oxidizers, we attempted enrichment of acidophilic archaeal ammonia oxidizers by inoculation of several acidic soils into mineral salts medium containing inorganic ammonium and adjusted to pH 4.5. An ammonia oxidizer enrichment culture was successfully obtained from an agricultural soil that had been maintained at pH 4.5 since 1961 and in which ammonia oxidation appears to be dominated by thaumarchaea (14). The enrichment contained a large number and diversity of heterotrophic "contaminants," and attempts were made to obtain a pure culture by successive subculturing in mineral salts medium, inclusion of streptomycin, and filtration. This led to a highly enriched culture of a thaumarchaeal ammonia oxidizer that is stable, in terms of its physiology and the coenriched bacterial community composition. A pure archaeal culture was also obtained by filtration, but failed to grow in the absence of the co-cultured bacteria, and addition of pyruvate did not enable the growth of the organism in pure culture, as reported recently for the cultivation of a neutrophilic ammonia oxidizer from soil (18). The cultivated archaeal strain grew exponentially in liquid batch culture, as demonstrated by coordinate, exponential increases in nitrite concentration and thaumarchaeal *amoA* gene abundance, assessed by quantitative PCR (qPCR), and by a decrease in ammonia concentration (Fig. 1). Cell activity and yield were estimated as $11 \text{ fmol-cell}^{-1}\text{-h}^{-1}$ and $4 \times 10^5 \text{ cells-}\mu\text{M ammonia}^{-1}$ (assuming that all nitrite was derived from ammonia oxidation), compared with respective values of $0.53 \text{ fmol-cell}^{-1}\text{-h}^{-1}$ and $4 \times 10^5 \text{ cells-}\mu\text{M ammonia}^{-1}$ reported for *Nitrosopumilus maritimus* (10).

Maximum *amoA* gene abundance was $1.3 \pm 0.47 \times 10^7$, and nitrite concentration did not increase to more than $41 \mu\text{M}$, de-

Author contributions: L.E.L.-M., J.I.P., and G.W.N. designed research; L.E.L.-M., K.S., and G.W.N. performed research; A.V., J.I.P., and G.W.N. contributed new reagents/analytic tools; L.E.L.-M., J.I.P., and G.W.N. analyzed data; and L.E.L.-M., J.I.P., and G.W.N. wrote the paper.

The authors declare no conflict of interest.

This article is a PNAS Direct Submission.

Data deposition: The sequences reported in this paper have been deposited in the GenBank database (accession nos. [JN227488](https://doi.org/10.1093/seqs/jn227488) and [JN227489](https://doi.org/10.1093/seqs/jn227489)).

¹To whom correspondence should be addressed. E-mail: j.prosser@abdn.ac.uk.

This article contains supporting information online at www.pnas.org/lookup/suppl/doi:10.1073/pnas.1107196108/-DCSupplemental.

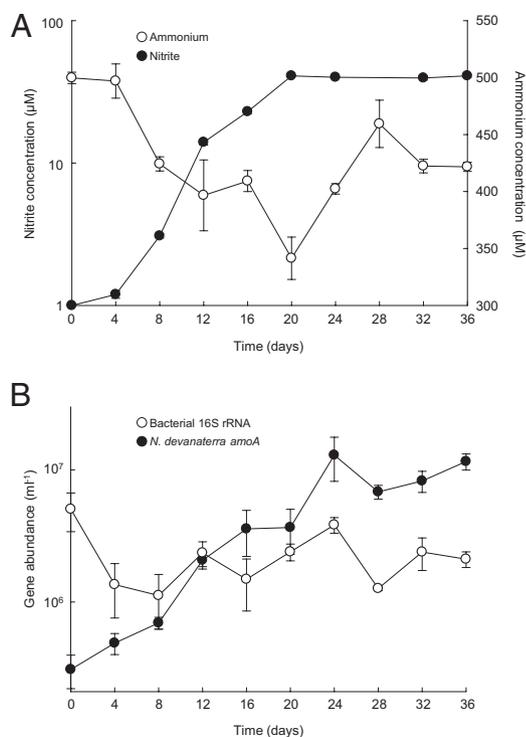


Fig. 1. Growth of, and ammonia oxidation by, an acidophilic thaumarchaeon in inorganic liquid medium at pH 4.5 containing 500 μM ammonium. (A) Exponential increase in nitrite concentration and accompanying decrease in ammonia concentration. (B) Exponential increase in thaumarchaeal *amoA* gene abundance, and growth of coenriched bacteria. Data are presented as mean and SE of triplicate cultures.

spite supply of 500 μM ammonium. Nitrite production and increases in thaumarchaeal *amoA* abundance were completely inhibited in the presence of 0.01% (10 Pa) acetylene (Fig. S1), which is a suicide substrate for ammonia monooxygenase (19) that is commonly used to inhibit autotrophic ammonia oxidation. qPCR provided no evidence of the presence of bacterial ammonia oxidizers (i.e., no amplification of 16S rRNA or *amoA* genes) or the presence of known nitrite oxidizers.

Ammonia concentration decreased during growth of the enriched culture, but conversion to nitrite was not stoichiometric. This is presumably because of the activity of coenriched bacteria, which could have used ammonia for growth and/or generated ammonium by mineralization of organic nitrogen carried over with the inoculum. The relative abundance of thaumarchaeal and bacterial 16S rRNA genes varied during growth, with bacterial abundance decreasing before thaumarchaeal growth began and remaining constant thereafter (Fig. 1B); bacteria presumably used organic carbon from the inoculum. The bacterial community contained organisms closely related to cultivated strains of *Burkholderia*, *Cupriavidus*, *Bradyrhizobium*, *Rhizobium*, *Mesorhizobium*, and *Elusimicrobium*, most of which have been isolated from root nodules and/or have the potential for nitrogen fixation (Table S1).

Assuming one *amoA* gene copy per thaumarchaeal genome (20, 21) and an estimated average of at least three 16S rRNA genes per bacterial genome (Table S1), qPCR analysis indicates that the thaumarchaeon represented approximately 90% of the cell abundance present during stationary phase. SEM of independent stationary-phase cultures indicated even greater dominance by small rod-shaped cells ($0.33 \pm 0.01 \mu\text{m}$ wide and $0.89 \pm 0.05 \mu\text{m}$ long) with a central indentation and electron-dense poles (Fig. 2D). These cells are morphologically very similar to the thaumarchaeon *N. maritimus* (22) and are clearly distinguishable from larger bacterial cells that were observed in

SEM images of early growth phase cultures, but very rare in SEM images from stationary phase. To confirm the identity of the small rod-shaped cells, FISH analyses were performed by using archaeal and bacterial probes ARCH915 and EUB338, respectively (Fig. 2A and B). The thaumarchaeon dominated cultures and represented more than 99% of all cells entering stationary phase. As observed previously for rod-shaped thaumarchaeal cells (22), ribosomes were largely concentrated at the poles, with SYBR Green 1 counterstaining revealing localization of the genome at the center and to one side of the cell (Fig. 2C).

Phylogenetic Analysis and Physiological Characteristics of Enriched Soil Thaumarchaeon. Phylogenetic analysis of 16S rRNA and *amoA* genes amplified from the enriched ammonia oxidizer places it within a group 1 lineage (Fig. 3) distinct from other cultivated thaumarchaea, with 89.8% and 77.4% identity to the 16S rRNA and *amoA* genes of *N. maritimus*, respectively. It does, however, have a relatively deep-branching association with group 1.1a organisms, forming a well supported monophyletic lineage distinct from group 1.1b and the *Nitrosocaldus yellowstonii* lineage, and is therefore described here as group 1.1a-associated (Fig. 3). The *amoBCxA* gene order was identical to that in *N. maritimus*, and a gene encoding 4-hydroxybutyryl-CoA dehydratase (*hcd*), a key enzyme and diagnostic of archaeal autotrophy (23), was also present.

The enrichment grew within the temperature range 20 to 30 $^{\circ}\text{C}$, with highest maximum specific growth rate (0.37 d^{-1}) at 25 $^{\circ}\text{C}$ (Fig. S2). No or little nitrite was detected after incubation for 28 d at 35 $^{\circ}\text{C}$ or at 15 $^{\circ}\text{C}$, despite enrichment from a temperate soil. The influence of pH was investigated by batch growth of the enriched thaumarchaeal ammonia oxidizer in liquid batch culture with initial pH adjusted to 3.5 to 7. Growth was restricted to the pH range 4.0 to 5.5 (Fig. 4A) when supplied with 500 μM ammonium, with no detectable changes in nitrite concentration or *amoA* abundance at pH 3.5 or at pH values greater than 5.5, and the organism may therefore be considered an obligate acidophile. Growth was optimal in the pH range of 4.0 to 5.0, with a maximum specific growth rate of $0.27 \pm 0.012 \text{ d}^{-1}$ or $0.23 \pm 0.01 \text{ d}^{-1}$, calculated from exponential increases in 16S rRNA gene abundance and nitrite concentration, respectively. These values are approximately 40% of the reported specific growth rate of 0.65 d^{-1} reported for *N. maritimus* (10). Maximum thaumarchaeal 16S rRNA gene abundance and nitrite concentration were observed at pH 5.0 and the pH response curve was shallow, with specific growth rates greater than 80% of the maximum. Reduced pH is believed to limit growth by reducing ammonia concentration and increasing nitrous/nitric oxide/ NO_x toxicity (10, 12). To assess the former, the thaumarchaeal ammonia oxidizer was grown with

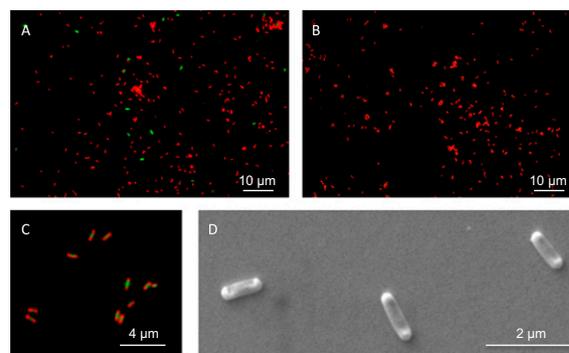


Fig. 2. Morphology and dominance of *N. devanattera* in an acidic ammonia oxidizing culture. FISH images with doubly labeled oligonucleotide probes at (A) midexponential and (B) onset of stationary phase growth of *N. devanattera*. Cells were cohybridized with Cy3-labeled ARCH915 (red) and FLUOS-labeled EUB338 (green) probes. (C) Cells probed with ARCH915 (red) and counterstained with DNA stain SYBR green 1. (D) SEM image of *N. devanattera* cells.

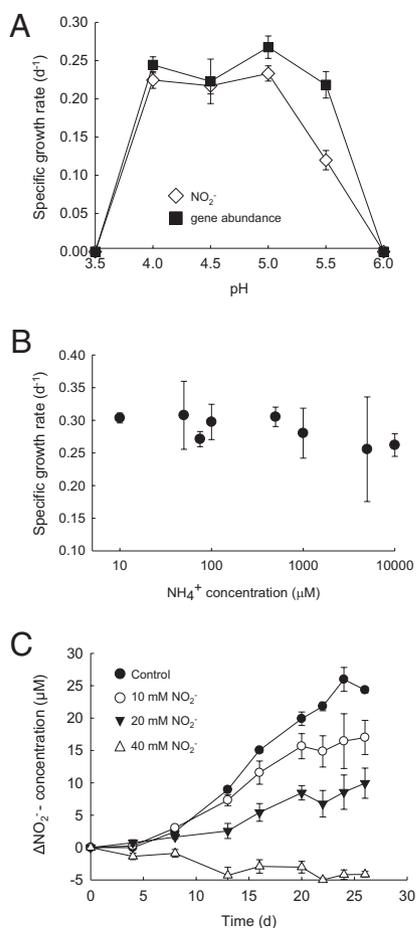


Fig. 4. The influence of pH and initial ammonium and nitrite concentrations on thaumarchaeal ammonia oxidizer growth. (A) Specific growth rate [from log-linear plots of nitrite concentration (diamonds) or thaumarchaeal 16S rRNA gene abundance (squares)] during exponential growth in medium containing 500 μM ammonium, initial pH 3.5 to 7.5. (B) Specific growth rate (from log-linear plots of nitrite concentration) during growth in medium containing initial ammonium concentration in the range of 10 μM to 10 mM, pH 4.5. (C) Changes in nitrite concentration during growth in medium containing 500 μM ammonium and initial nitrite concentration in the range of 10 to 40 μM, pH 4.5. Data are presented as mean and SE of triplicate cultures for all growth experiments.

autotrophic growth of this phylotype. Comparison of the distribution in CsCl gradients of genomic DNA extracted from microcosms incubated with 98% ¹²C₂O₂ or 99% ¹³C₂O₂ demonstrated an increase in the buoyant density of *N. devanaterrea*-like genomic DNA from approximately 1.69 g·mL⁻¹ to 1.73 g·mL⁻¹ (Fig. 5D). Although this approach lacks precision, a difference of 0.04 g·mL⁻¹ would be expected between unlabeled ¹²C- and fully labeled ¹³C-DNA (24), thus indicating that *N. devanaterrea* grows in soil using mainly inorganic carbon.

Discussion

Previous explanations for the common observation of relatively high rates of nitrification in soil, despite only neutrophilic growth of ammonia oxidizer laboratory isolates in standard laboratory culture, have been difficult to demonstrate in soil. For example, many bacterial ammonia oxidizers have ureolytic activity (25), which enables growth on urea at low pH (7, 8), but urea will derive mainly from urine patches in grazed land and is usually converted rapidly to ammonia by extracellular soil enzymes and ureolytic heterotrophs. Ureolytic activity may therefore enable ammonia oxidation only in acid soils under restricted conditions. Other

explanations involve growth in biofilms or in multicellular aggregates, and acidophilic growth in such microenvironments is difficult to demonstrate unequivocally. In addition, these explanations are based on physiological characteristics of bacterial ammonia oxidizers, which are often undetectable in acid soils (15, 26, 27), and predate demonstration of the capacity for ammonia oxidation by thaumarchaea in soil (28). Here we report the discovery of an obligately acidophilic ammonia oxidizer, *Candidatus Nitrosotalea devanaterrea*, that does not grow at neutral pH, but whose growth and activity at low pH are similar to those of soil bacterial ammonia oxidizer enrichments and isolates at neutral pH. Phylogenetic analysis of 16S rRNA and *amoA* genes places the organism within the group 1.1a-associated thaumarchaea and we propose the following candidate status:

“*Nitrosotalea devanaterrea*” gen. et sp. nov.

Etymology. Nitrosus (Latin masculine adjective), nitrous; talea (Latin feminine noun), slender rod; devana (Latin), Aberdeen; terra (Latin feminine noun), earth, soil. The name reflects its ability to oxidize ammonia to nitrite, its morphology and its site of origin.

Locality. Acidic agricultural soil, Craibstone, Aberdeen, Scotland, United Kingdom.

Diagnosis. A chemolithoautotrophic ammonia oxidizer of the kingdom *Thaumarchaeota*, appearing as straight rods with diameter 0.33 ± 0.01 μm and length 0.89 ± 0.05 μm.

The discovery of *Nitrosotalea devanaterrea* does not exclude a role for bacterial ammonia oxidizers in nitrification in acid soils, nor involvement of urease activity or biofilm and aggregate formation in enabling acidophilic nitrification. It does, however, provide the most parsimonious explanation for ammonia oxidation in acid soils, which is reinforced by the common occurrence of closely related phylotypes in acid soils. This is further reinforced by demonstration of autotrophic ammonia oxidation by closely related phylotypes in acid soil microcosms.

The current inability to eliminate heterotrophs from the *N. devanaterrea* enrichment culture limits the scope of physiological studies, but, in common with other ammonia oxidizers, activity is inhibited by acetylene in liquid and in soil. It is intriguing that many of the coenriched heterotrophs have the potential for nitrogen fixation or are associated with legume root nodules. This suggests a possible commensal or symbiotic association, particularly as their removal prevented growth. However, the enrichment cultivation conditions are unlikely to favor nitrogen fixation, their presence may be fortuitous, and additional enrichments are required to assess the significance of this finding.

Acidophilic ammonia oxidation suggests novel physiological mechanisms in *N. devanaterrea*. Its growth is consistent with the half-saturation constant (K_m) of 133 nM total ammonium reported for *N. maritimus* (29), but not if measured in terms of ammonia, which is generally assumed to be the substrate for ammonia monooxygenase. Growth occurred at pH 4.5 and 0.18 nM ammonia, and specific growth rate decreased at higher ammonia concentrations, suggesting that this organism has developed potentially novel system(s) for ammonia oxidation, in addition to protective mechanisms enabling acidophilic growth. Further characterization of mechanisms for acidophilic growth and ammonia oxidation are of considerable interest in terms of the physiology and evolution of this lineage, and its discovery may even require reassessment of the paradigm of ammonia, rather than ammonium, being the substrate for the membrane-bound ammonia monooxygenase. Of particular interest will be the location of the active site, i.e., whether it faces outwards into a periplasmic-like space, where it may be significantly influenced by the pH of the external environment.

Growth of *N. devanaterrea* was inhibited by 40 μM nitrite, which, at pH 4.5, is equivalent to 2.53 μM free nitrous acid concentration (HNO₂), which is believed to be responsible for inhibition of ammonia oxidation. This concentration is lower than reported K_i values, which lie within the range of 12 to 200 μM HNO₂ (30, 31),

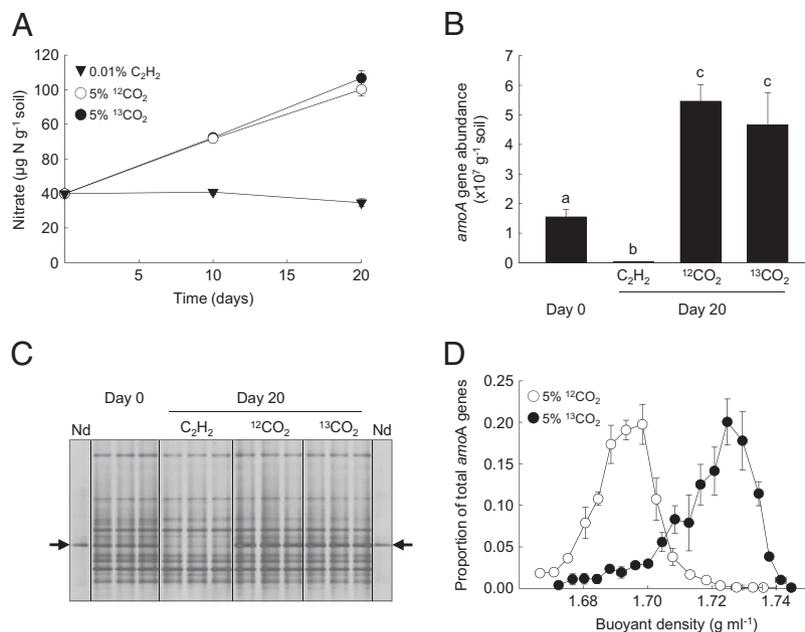


Fig. 5. Autotrophic growth and ammonia oxidation by *N. devanattera* phylotypes in an acidic soil. Triplicate microcosms were incubated with 5% ^{12}C - or ^{13}C - CO_2 or 0.01% acetylene for 0, 10, or 20 d. (A) Changes in nitrate concentration. (B) Abundance of group 1.1a-associated *amoA* genes at 0 and 20 d. Different letters indicate statistically significant differences ($P < 0.05$). (C) DGGE analysis of thaumarchaeal *amoA* PCR products from soil and from *N. devanattera* (Nd). Arrows indicate the migration of *N. devanattera* phylotypes. Each lane represents an individual microcosm. (D) Distribution of group 1.1a-associated thaumarchaeal genomic DNA after separation and fractionation in CsCl gradients. Vertical and horizontal bars represent SEs of proportional abundance and buoyant density of fractions, respectively, derived from triplicate CsCl gradients representing triplicate microcosms.

and may indicate greater sensitivity to nitrous acid. Regardless of the mechanism of inhibition, continued nitrification would require close interactions with acidophilic nitrite oxidizers to relieve nitrous acid-associated inhibition in soil. Coculture with an acidophilic nitrite oxidizer, or other means of removing nitrite, might also enable more extended growth of *N. devanattera* (10).

In conclusion, enrichment of an autotrophic, obligately acidophilic ammonia oxidizing thaumarchaeon, commonly found in acid soils, and demonstration of its autotrophic growth in acid soil provide a new and convincing solution to the paradox of nitrification in acid soils. It also suggests novel mechanisms for activity and growth of ammonia oxidizers and, in particular, challenges current paradigms surrounding substrate uptake mechanisms.

Materials and Methods

Cultivation and Growth Experiments. Ammonia oxidizer enrichment cultures were established in 100-mL Duran bottles containing 50 mL of an unbuffered mineral salts medium based on that of Tourna et al. (18) consisting of NaCl (1 g·L⁻¹), MgCl₂ (0.4 g·L⁻¹), CaCl₂ (0.1 g·L⁻¹), KH₂PO₄ (0.2 g·L⁻¹), KCl (0.5 g·L⁻¹), 1 mL modified nonchelated trace element solution (32), 1 mL 7.5 mM NaFeEDTA, 2 mM NaHCO₃, 500 µM NH₄Cl and 50 mg·L⁻¹ streptomycin. The pH of the medium was adjusted to 4.5 with HCl, before filter-sterilization through a 0.22 µm pore size, bottle-top filter. Medium was inoculated (1%, wt/vol) with an agricultural sandy loam soil (Scottish Agricultural College; grid reference NJ872104) (13). Cultures were incubated at 28 °C in the dark and ammonia oxidizer activity and growth were assessed by analysis of inorganic nitrogen and thaumarchaeal 16S rRNA and *amoA* gene abundance, respectively. All physiology experiments were conducted in triplicate cultures. The effect of ammonia concentration was investigated in medium containing 10, 50, 100, and 500 µM and 1, 5, 10, 50, and 100 mM NH₄Cl. Inhibition of growth by nitrite was studied in medium containing initial nitrite (i.e., NaNO₂) concentrations of 0, 10, 20, and 40 µM. The effects of pH and temperature were investigated by incubating cultures at pH values in the range 3.5 to 7.0 (0.5 pH intervals) and temperatures in the range of 15 to 40 °C (5 °C intervals), respectively. Incubations with acetylene were performed in 144-mL serum vial bottles containing 50 mL of culture. Bottles were crimp-sealed, and 0.01% acetylene (vol/vol) headspace concentration was established. Bottles were opened at 4-d intervals for sampling and reestablishment of acetylene concentration. Specific growth rate was cal-

culated from log-linear plots of gene abundance or nitrite concentration for individual cultures under a variety of treatments.

Process Measurements. Nitrite and ammonia concentrations were determined colorimetrically (33, 34). Standards consisted of NaNO₂ (0.781–50 µM) or NH₄Cl (15.6–500 µM) prepared in mineral salts medium. All reactions and absorbance measurements were performed in clear polystyrene 96-well microplates (Greiner Bio-One) by using a LT-4000 Microplate Reader (Labtech).

qPCR. Thaumarchaeal growth was assessed by changes in abundance of 16S rRNA and *amoA* genes. DNA was extracted by using a bead-beating protocol with SDS-based buffer, phenol, chloroform, and isoamyl alcohol (35). DNA precipitation was performed as described previously (35) but with the inclusion of linear acrylamide. qPCR amplification was performed by using a MyIQ real-time PCR detection system (BioRad) with QuantiFast SYBR Green 1 PCR Master Mix (Qiagen). qPCR was performed as previously described (36) by using thaumarchaeal-specific 16S rRNA gene primers 771F and 957R (37) and group 1.1a-associated/*N. devanattera* primers 167F (5'-TATCAATCA-YTGATGCTCGCAGT-3') and 409R (5'-TCATGTTGAACAWCGACAT-3') designed in this study by using group 1.1a-associated sequences, and discriminating against other group 1 lineages. qPCR standards for thaumarchaeal 16S rRNA gene and group 1.1a-associated *amoA* gene assays were a cloned 1.3-kb thaumarchaeal gene sequence and *N. devanattera amoBCxA* amplicons, respectively. Bacterial qPCR was performed as described previously (36). qPCR efficiencies for the thaumarchaeal 16S rRNA gene, group 1.1a-associated *amoA* gene, and bacterial 16S rRNA gene assays were 97.0 to 105.1%, 91.1 to 101.2%, and 98.2 to 101.0%, respectively, with r^2 values of at least 0.99 for all assays.

Sequence Analyses. Thaumarchaeal 16S rRNA genes were amplified from enrichment culture DNA by using primers A109F (38) and 1492R (13) generating a product approximately 1.3 kb in length. The 3-kb *amo* gene cluster (containing *amoB*, *amoC*, a conserved *amo*-associated hypothetical ORF, and *amoA*) was amplified by using primers 616R and *amo2.2R* (13, 22), and entire 16S rRNA and *amoA* gene amplicons were sequenced along both strands. The *hcd* gene was amplified as described previously (23). Phylogenetic analyses were performed on both DNA and inferred amino acid sequences using distance (39), parsimony (39), maximum-likelihood (40), and Bayesian methods (41) with GTR (DNA), JTT, or mixed model (amino acid) correction with ML-estimated (40) γ -distributed variable sites. To identify cocultured

bacteria, 16S rRNA genes were amplified by using primers 27F and 1492R (42) from DNA extracted from three independent cultures before cloning and transforming into competent *Escherichia coli* by using standard techniques. Clones were randomly selected from each library and sequenced in one direction, and phylogenetic affiliation inferred through BLASTN searches. DGGE analysis was performed on total thaumarchaeal *amoA* gene amplicons as described previously (13).

SEM. Glutaraldehyde-fixed cell suspensions from a stationary culture were applied to glass coverslips coated with poly-L-lysine, allowed to adhere for 5 min, rinsed with water, and postfixed with 1% OsO₄. Attached cells were then rinsed with distilled H₂O, dehydrated through a graded ethanol series to absolute ethanol, and critically point-dried with liquid CO₂. Coverslips were sputtered with Au before examination by using an Evo MA 10 scanning electron microscope (Carl Zeiss).

DOPE-FISH. Cultures were fixed with paraformaldehyde, and FISH was performed by using doubly labeled oligonucleotide probes (43) using Cy3 doubly labeled probe Arch915 and the FLUOS doubly labeled probe EUB338, respectively. Counterstaining with the nucleic acid-binding dye SYBR Green I revealed a detection rate greater than 99%. Hybridizations were analyzed using a Leica DM500 microscope equipped with a DFC345 FX Peltier-cooled CCD camera (Wetzlar).

1. von Uexküll HR, Mutert E (1995) *Plant-Soil Interactions at Low pH, Principles and Management*, eds Date RA, Grundon NJ, Raymet GE, Probert ME (Kluwer, New York), pp 5–19.
2. Booth MS, Stark JM, Rastetter E (2005) Controls on nitrogen cycling in terrestrial ecosystems: A synthetic analysis of literature data. *Ecol Monogr* 75:139–157.
3. Allison SM, Prosser JI (1991) Urease activity in neutrophilic autotrophic ammonia-oxidizing bacteria isolated from acid soils. *Soil Biol Biochem* 23:45–51.
4. Jiang QQ, Bakken LR (1999) Comparison of *Nitrosospora* strains isolated from terrestrial environments. *FEMS Microbiol Ecol* 30:171–186.
5. Suzuki I, Dular U, Kwok SC (1974) Ammonia or ammonium ion as substrate for oxidation by *Nitrosomonas europaea* cells and extracts. *J Bacteriol* 120:556–558.
6. Frijlink MJ, Abee T, Laanbroek HJ, de Boer W, Konings WN (1992) The bioenergetics of ammonia and hydroxylamine oxidation in *Nitrosomonas europaea* at acid and alkaline pH. *Arch Microbiol* 157:194–199.
7. de Boer W, Duyts H, Laanbroek HJ (1989) Urea stimulated autotrophic nitrification in suspensions of fertilized, acid heath soil. *Soil Biol Biochem* 21:349–354.
8. Burton SAQ, Prosser JI (2001) Autotrophic ammonia oxidation at low pH through urea hydrolysis. *Appl Environ Microbiol* 67:2952–2957.
9. Killham K (1986) Heterotrophic nitrification. *Nitrification*, ed Prosser JI (IRL Press, Oxford), pp 117–126.
10. De Boer W, Gunnewiek PJ, Veenhuis M, Bock E, Laanbroek HJ (1991) Nitrification at low pH by aggregated chemolithotrophic bacteria. *Appl Environ Microbiol* 57:3600–3604.
11. Allison SM, Prosser JI (1993) Ammonia oxidation at low pH by attached populations of nitrifying bacteria. *Soil Biol Biochem* 25:935–941.
12. de Boer W, Kowalchuk GA (2001) Nitrification in acid soils: Micro-organisms and mechanisms. *Soil Biol Biochem* 33:853–866.
13. Nicol GW, Leininger S, Schleper C, Prosser JI (2008) The influence of soil pH on the diversity, abundance and transcriptional activity of ammonia oxidizing archaea and bacteria. *Environ Microbiol* 10:2966–2978.
14. Gubry-Rangin C, Nicol GW, Prosser JI (2010) Archaea rather than bacteria control nitrification in two agricultural acidic soils. *FEMS Microbiol Ecol* 74:566–574.
15. Stopnišek N, et al. (2010) Thaumarchaeal ammonia oxidation in an acidic forest peat soil is not influenced by ammonium amendment. *Appl Environ Microbiol* 76:7626–7634.
16. Leininger S, et al. (2006) Archaea predominate among ammonia-oxidizing prokaryotes in soils. *Nature* 442:806–809.
17. He JZ, et al. (2007) Quantitative analyses of the abundance and composition of ammonia-oxidizing bacteria and ammonia-oxidizing archaea of a Chinese upland red soil under long-term fertilization practices. *Environ Microbiol* 9:2364–2374.
18. Tournia M, et al. (2011) *Nitrososphaera viennensis*, an ammonia oxidizing archaeon from soil. *Proc Natl Acad Sci USA* 108:8420–8425.
19. Hyman MR, Wood PM (1985) Suicidal inactivation and labelling of ammonia monooxygenase by acetylene. *Biochem J* 227:719–725.
20. Walker CB, et al. (2010) *Nitrosopumilus maritimus* genome reveals unique mechanisms for nitrification and autotrophy in globally distributed marine crenarchaea. *Proc Natl Acad Sci USA* 107:8818–8823.
21. Hallam SJ, et al. (2006) Pathways of carbon assimilation and ammonia oxidation suggested by environmental genomic analyses of marine *Crenarchaeota*. *PLoS Biol* 4:e95.
22. Könneke M, et al. (2005) Isolation of an autotrophic ammonia-oxidizing marine archaeon. *Nature* 437:543–546.
23. Offre PO, Nicol GW, Prosser JI (2010) Autotrophic community profiling and quantification of putative autotrophic thaumarchaeal communities in environmental samples. *Environ Microbiol Rep* 3:245–253.
24. Buckley DH, Huangyutitham V, Hsu S-F, Nelson TA (2007) Stable isotope probing with ¹⁵N achieved by disentangling the effects of genome G+C content and isotope enrichment on DNA density. *Appl Environ Microbiol* 73:3189–3195.
25. Koops H-P, Böttcher B, Möller UC, Pommerening-Röser A, Stehr G (1991) Classification of eight new species of ammonia-oxidising bacteria: *Nitrosomonas communis* sp. nov., *Nitrosomonas ureae* sp. nov., *Nitrosomonas aestuarii* sp. nov., *Nitrosomonas marina* sp. nov., *Nitrosomonas nitrosa* sp. nov., *Nitrosomonas eutropha* sp. nov., *Nitrosomonas oligotropha* sp. nov. and *Nitrosomonas halophila* sp. nov. *J Gen Microbiol* 137:1689–1699.
26. Klemmedtsson L, Jiang Q, Klemmedtsson AK, Bakken L (1999) Autotrophic ammonium-oxidising bacteria in Swedish mor humus. *Soil Biol Biochem* 31:839–847.
27. Schmidt CS, et al. (2007) PCR profiling of ammonia-oxidizer communities in acidic soils subjected to nitrogen and sulphur deposition. *FEMS Microbiol Ecol* 61:305–316.
28. Zhang L-M, et al. (2010) Autotrophic ammonia oxidation by soil thaumarchaea. *Proc Natl Acad Sci USA* 107:17240–17245.
29. Martens-Habbena W, Berube PM, Urakawa H, de la Torre JR, Stahl DA (2009) Ammonia oxidation kinetics determine niche separation of nitrifying Archaea and Bacteria. *Nature* 461:976–979.
30. Park S, Bae W (2009) Modeling kinetics of ammonium oxidation and nitrite oxidation under simultaneous inhibition by free ammonia and free nitrous acid. *Process Biochem* 44:631–640.
31. Anthonisen AC, Loehr RC, Prakasam TBS, Srinath EG (1976) Inhibition of nitrification by ammonia and nitrous acid. *J Water Pollut Control Fed* 48:835–852.
32. Widdel F, Bak F (1992) The Prokaryotes. *A Handbook on the Biology of Bacteria: Ecophysiology, Isolation, Identification, Application*, eds Ballows A, Trüper HG, Dworkin M, Harder W, Schleifer K-H (Springer, New York), 2nd Ed, pp 3352–3378.
33. Shinn MB (1941) Colorimetric method for determination of nitrite. *Ind Eng Chem Anal Ed* 13:33–35.
34. Crooke WM, Simpson WE (1971) Determination of ammonium in Kjeldahl digests of crops by an automated procedure. *J Sci Food Agric* 22:9–10.
35. Tournia M, Freitag TE, Nicol GW, Prosser JI (2010) Stable isotope probing analysis of interactions between ammonia oxidizers. *Appl Environ Microbiol* 76:2468–2477.
36. Lehtovirta LE, Prosser JI, Nicol GW (2009) Soil pH regulates the abundance and diversity of Group 1.1c Crenarchaeota. *FEMS Microbiol Ecol* 70:367–376.
37. Ochsenreiter T, Selez D, Quaiser A, Bonch-Osmolovskaya L, Schleper C (2003) Diversity and abundance of Crenarchaeota in terrestrial habitats studied by 16S RNA surveys and real time PCR. *Environ Microbiol* 5:787–797.
38. Grosskopf R, Janssen PH, Liesack W (1998) Diversity and structure of the methanogenic community in anoxic rice paddy soil microcosms as examined by cultivation and direct 16S rDNA gene sequence retrieval. *Appl Environ Microbiol* 64:960–969.
39. Tamura K, Dudley J, Nei M, Kumar S (2007) MEGA4: Molecular evolutionary genetics analysis (MEGA) software version 4.0. *Mol Biol Evol* 24:1596–1599.
40. Guindon S, Gascuel O (2003) A simple, fast, and accurate algorithm to estimate large phylogenies by maximum likelihood. *Syst Biol* 52:696–704.
41. Ronquist F, Huelsenbeck JP (2003) MrBayes 3: Bayesian phylogenetic inference under mixed models. *Bioinformatics* 19:1572–1574.
42. Lane DJ (1991) *Nucleic Acid Techniques in Bacterial Systematics*, eds Stackebrandt E, Goodfellow M (Wiley, New York), pp 115–175.
43. Stoecker K, Dorninger C, Daims H, Wagner M (2010) Double labeling of oligonucleotide probes for fluorescence in situ hybridization (DOPE-FISH) improves signal intensity and increases rRNA accessibility. *Appl Environ Microbiol* 76:922–926.