

The Importance of Islands for the Protection of Biological and Linguistic Diversity

BERNIE R. TERSHY, KUO-WEI SHEN, KELLY M. NEWTON, NICK D. HOLMES, AND DONALD A. CROLL

Islands make up 5.3% of Earth's land area yet maintain an estimated 19% of bird species, 17% of rodents, 17% of flowering plants, and 27% of human languages. Species diversity is disproportionately threatened on islands in relation to the islands' proportion of both global land area and species, with 61% of all extinct species and 37% of all critically endangered species confined to islands. Languages are disproportionately threatened on islands in relation to land area with 11% of extinct languages and 25% of critically endangered languages on islands. Islands are a priority area for integrated conservation efforts because they have 14 times greater density of critically endangered terrestrial species and 6 times greater density of critically endangered languages than continental areas. Invasive species and habitat loss are the largest threats to island terrestrial species diversity. Proven management actions can reduce these threats, benefiting both local peoples and species diversity on islands.

Keywords: biodiversity, extinction, linguistics, islands, conservation

We are in the middle of an extinction crisis brought about by land conversion, overexploitation, pollution, and invasive species (Pimm et al. 2006). For well-studied taxa, current extinction rates are three orders of magnitude greater than background rates and equally above rates at which new species evolve (Pimm et al. 2014). This loss of species has negative economic, ethical, and aesthetic impacts and is permanent over time scales relevant to humans. Within our own species, there is a concurrent loss of linguistic diversity (Maffi 2005), with nearly 3% of known languages lost in the last three generations (Lewis et al. 2013). This loss of linguistic diversity has cultural, ethical, and scientific impacts and is also permanent over time scales relevant to humans (Austin and Sallabank 2011).

Efforts to prevent the ongoing loss of biological and linguistic diversity have been extensive. For biodiversity in particular, there have been a number of approaches that prioritize conservation efforts in areas where concentrations of all species—or of threatened species—are highest (e.g., Myers et al. 2000, Brooks et al. 2006). Several authors have demonstrated the synergy between biological and linguistic diversity (Maffi 2007). Others have advocated for the inclusion of linguistic diversity with biodiversity hotspots, wilderness areas, and other biogeographical parameters, including the comparison of their distribution on islands with that in mainland areas (Gorenflo et al. 2012). Islands warrant a unique level of attention for biodiversity conservation

because they make up only a small percentage of land area but are known for their many endemic species (Kier et al. 2009, Weigelt et al. 2013). Here, we present a detailed examination of the concentration of species diversity and linguistic diversity on islands. Specifically, we quantify the number and density of total, critically endangered, and extinct species and human languages on islands and continents and the causes of species endangerment and extinction on islands and continents. We conclude by suggesting potentially synergistic efforts for the conservation of these two types of diversity.

Methods

We used the Global Island Database to determine the number (greater than 180,000) and area (7,820,560 square kilometers [km²]) of islands on Earth (UNEP-WCMC 2013). The Global Island Database lists all islands larger than 0.11 hectares (ha). Estimates of the percentage of global species diversity on islands were from published sources and only available for flowering plants (17%; Whittaker and Fernández-Palacios 2007), birds (19%; Newton 2003), and rodents (17%; Amori et al. 2008).

Data on the threat status of island species for all taxa of plants and animals are from the International Union for Conservation of Nature (IUCN) Red List of Threatened Species (version 2010.1; IUCN 2010). For each species listed by the IUCN as *extinct* (including *extinct in the*

Table 1. The number of species and languages in each category and their threat status.

| | | Insular | Continental | Total | Percentage insular |
|---|-----------------|---------|-------------|---------|--------------------|
| Area (in 1000 square kilometers) | | 7821 | 141,118 | 140,939 | 5.3 |
| Total diversity | Magnoliophyta | 48,331 | 241,669 | 290,000 | 17 |
| | Aves | 1947 | 8117 | 10,064 | 19 |
| | Rodentia | 388 | 1847 | 2262 | 17 |
| | Languages | 2551 | 6885 | 9436 | 27 |
| Extinct diversity | Plantae | 71 | 43 | 114 | 62.3 |
| | Animalia total | 461 | 296 | 757 | 60.9 |
| | Actinopterygii | 4 | 99 | 103 | 3.9 |
| | Amphibian | 21 | 18 | 39 | 53.8 |
| | Arthropoda | 38 | 31 | 69 | 55.1 |
| | Aves | 127 | 10 | 137 | 95.3 |
| | Mammalia | 42 | 36 | 78 | 53.8 |
| | Mollusca | 212 | 97 | 309 | 68.6 |
| | Platyhelminthes | 0 | 1 | 1 | 0 |
| | Reptilia | 17 | 4 | 21 | 81.0 |
| | Languages | 43 | 336 | 379 | 11.4 |
| Critically endangered diversity | Plantae | 794 | 781 | 1575 | 50.4 |
| | Animalia total | 641 | 1100 | 1741 | 36.8 |
| | Actinopterygii | 49 | 240 | 289 | 16.9 |
| | Amphibian | 102 | 382 | 484 | 21.1 |
| | Arthropoda | 94 | 85 | 179 | 52.5 |
| | Aves | 113 | 79 | 192 | 58.8 |
| | Chondrichthyes | 0 | 25 | 25 | 0 |
| | Mammalia | 85 | 103 | 188 | 45.2 |
| | Mollusca | 154 | 137 | 291 | 52.9 |
| | Reptilia | 44 | 49 | 93 | 47.3 |
| | Languages | 113 | 347 | 460 | 24.6 |

Note: See the “Methods” section for references.

wild) and each species listed by the IUCN as *critically endangered*, we used distribution data from IUCN Red List maps, peer-reviewed literature, and additional sources to determine whether it was found exclusively on islands (if a species distribution spanned insular and continental habitats, it was considered a continental species). We then determined the relative severity of the main drivers of extinction—habitat loss, invasive species, overexploitation, pollution, and other causes, including climate change and disease—by tallying the number of times each threat was mentioned in the IUCN Red List species account as contributing to the extinction or endangerment of each species. We included disease in the *other* category because we were unable to consistently determine whether a disease was native or introduced. We included climate change in the *other* category rather than as a type of pollution because it was inconsistently applied in the IUCN Red List species account. Sometimes, a climate change threat was specific; other times, it was a generalized threat that could be applied to most species.

Data on linguistic diversity, distribution, and threat status are from Lewis (2009). Lewis (2009) provided a list of all known languages, and for each language, we used the accompanying geographic data to determine whether it was confined to islands or at least partially continental. To improve readability, we call his “nearly extinct” languages *critically endangered*. Data were not available on the causes of language extinction and endangerment.

Results

Islands make up 5.3% of Earth’s land area, but an estimated 17% of plant species, 19% bird species, and 17% of rodent species are confined to islands, as are 27% of human languages (table 1). Consequently, the density of species and language diversity on islands is higher than on continents (figure 1).

Sixty-one percent of all species listed by the IUCN as *extinct* and 37% of species listed by the IUCN as *critically endangered* are confined to islands (table 1). Therefore, the density of extinct and critically endangered species is

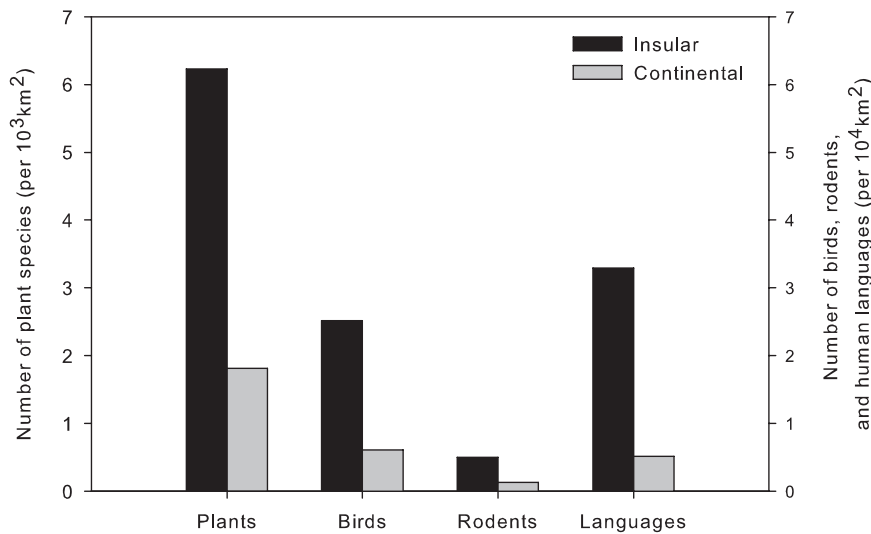


Figure 1. Number of insular and continental plant species per 10³ square kilometers (km²), bird and rodent species and human languages per 10⁴ km² confined to islands (insular) or occurring primarily, but not exclusively, on continental areas (continental).

much greater on islands than on continents in relation to that expected according to land area or number of species (figure 2). Linguistic diversity, in contrast, is somewhat less vulnerable on islands than would be expected according to the number of languages, with 11% of extinct languages and 25% of critically endangered languages confined to islands (table 1, figure 2).

On islands, invasive species were cited most frequently as the cause of species extinctions and were the second most frequently cited cause of critical endangerment (figure 3).

Conclusions

Marine islands are disproportionately rich in species and linguistic diversity. When corrected for surface area, islands have 3.6 more species per km² and 6.7 times more languages per km² than continental areas (figure 1). This is a somewhat conservative measure of the biological and linguistic importance of islands, because it only includes species and languages entirely confined to islands (insular endemics) and lumps species and languages that occur exclusively on continental areas with those that occur on both continental areas and islands. The inclusion of more taxonomic groups and new data on plant, bird, and rodent diversity confined to islands will undoubtedly change the details of our results. However, new data are unlikely to dramatically change the overall trend of islands having a density of unique terrestrial species diversity several times greater than continental areas. A similar conclusion was reached by Kier and colleagues (2009), who found that oceanic islands and Peninsular Malaysia had 9.5 times greater plant endemism richness and 8.1 times greater vertebrate endemism richness than continental areas.

Marine island species have been disproportionately affected by ongoing global changes. More than half of

the world's known extinct species and almost 40% of species listed by the IUCN as *critically endangered* are confined to islands. When we corrected for surface area, islands have almost 30 times more extinct species per km² and 14 times more critically endangered species per km² than continental areas (figure 2). Previous studies on quantifying the proportion of threatened and extinct species on islands included only better-studied threatened taxa, such as birds and mammals (Ricketts et al. 2005, Loehle and Eschenbach 2012). Our results for these taxa are generally similar to those, with only minor differences due to recent Red List updates. As more taxonomic groups (particularly plants and invertebrates) are assessed for the IUCN Red List, the percentage of extinct and threatened species confined to islands will also change. However, these new data are unlikely to

dramatically change the overall trend of island species being more threatened than mainland species.

In contrast to that of species diversity, the percentage of extinct languages on islands is lower than the percentage of languages confined to islands, and the percentage of endangered languages on islands is equal to the percentage of all languages on islands (table 1, figure 2). Although language is a product of the social and natural environment (Halliday 2001) and is therefore influenced by species losses and introductions, language extinctions are complex and more directly linked to interactions between cultures, the decline of human lineages, and the adoption of competing languages (Wurm 2003), which makes them challenging to compare with biological extinctions (Sutherland 2003). Geographically, whereas linguistic and biological diversity tend to show some correlation on islands around the globe, this relationship is not strong (Gorenflo et al. 2012). So although it is reasonable to expect the same general patterns for threatened species and languages, there will be important divergences. For example, the islands in East Melanesia are both linguistic and species diversity hotspots, whereas Madagascar is a species diversity hotspot but not a linguistic hotspot (which is an artifact of relatively recent human settlement). Determining which regions and islands of the world show the strongest overlap between threatened languages and species would provide valuable insight into developing synergistic conservation opportunities.

Insular species may be more vulnerable to extinction and endangerment than continental species, because they have smaller population sizes and smaller ranges (MacArthur and Wilson 1967) and less genetic diversity (Frankham 1997) and because they lack behavioral (Blumstein and Daniel 2005), life-history (Köhler and Moyà-Solà 2009), and morphological (Bowen and Vuren 1997) defenses

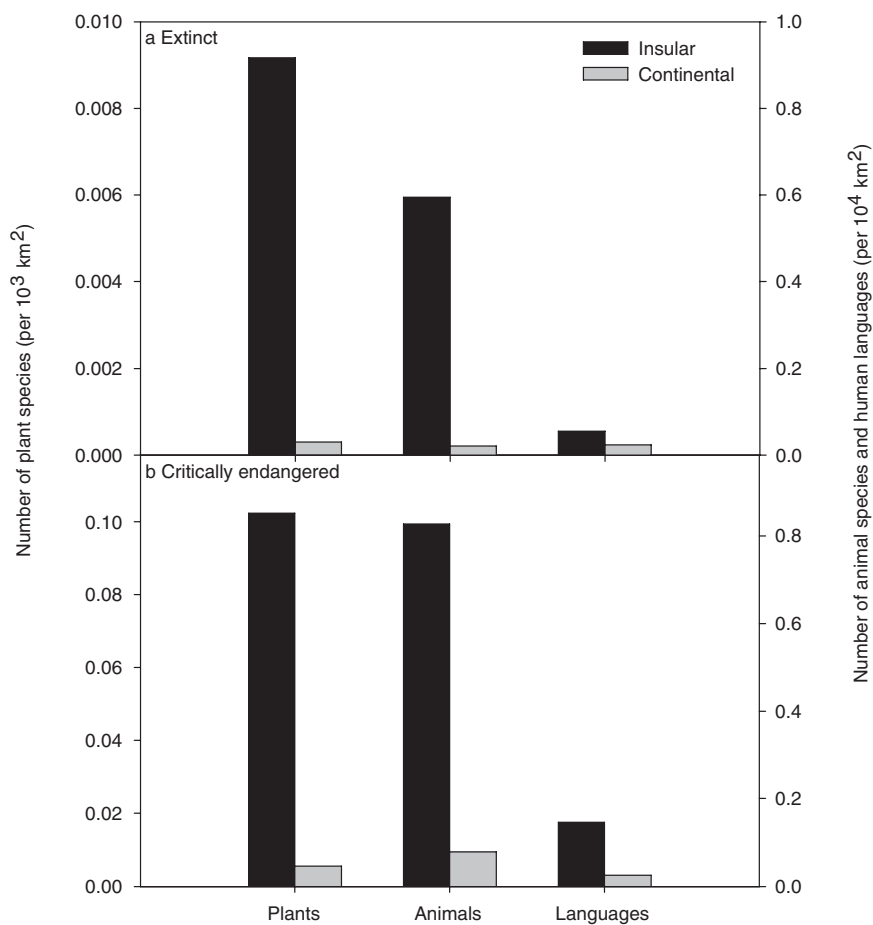


Figure 2. The number of extinct (a) and critically endangered (b) plant species (per 10³ square kilometers [km²]) and all animal species and languages (per 10⁴ km²) confined to islands (insular) or occurring primarily but not exclusively on continental areas (continental). The species data are from IUCN (2010); the language data are from Lewis (2009).

against human predators and invasive predators and herbivores which often occur at extremely high densities on islands (Terborgh et al. 2001). Of the four primary drivers of species loss, invasive species were the most frequently cited cause for the decline of extinct island species, and habitat loss was the most frequently cited cause for the decline of critically endangered island species (figure 3). The timing of these two threats occurring on islands likely contributes to this difference, with the introduction of invasive species rapidly leading to the early extinction of many vulnerable species (Hedges and Conn 2012) before habitat loss becomes a significant threat. It is clear that these two threats are ongoing drivers of extinction for species and that they can be exacerbated by pollution-induced climate change (Martin et al. 2012).

Although efforts to protect endangered languages are not necessarily the same as efforts to protect endangered species (Maffi 2007, Pretty et al. 2009), the high concentration of both types of diversity on islands suggests that some efforts to protect threatened island species

may also protect threatened island languages. Functionally, the downlisting of threatened species requires increasing population size or distribution, ideally within native habitat, whereas reversing language endangerment requires increasing the number of speakers and providing a healthy cultural environment for these languages to be authentically used. Protecting native species and ecosystems protects ecosystem services that can improve the resilience and livelihoods of the rural communities where most threatened languages are found (Arnold et al. 2011, Sangha et al. 2011, Kalaba et al. 2013). Furthermore, several avenues of linguistic research offer unique opportunities to improve conservation practice and maintain cultural identity for local indigenous populations. Involving traditional ecological knowledge (TEK) in endangered species management provides an opportunity to integrate protection of language, cultural practice, and native species, such as Maori TEK on tuatara-inhabited islands in New Zealand (Ramstead et al. 2007). Ethnobotanical research provides an important practice to improve our understanding of natural systems, including basic taxonomic description, ecological adaptation, and conservation status of species.

The isolation and simplified ecosystems of small islands may facilitate a suite of actions to protect both species

diversity and linguistic diversity—or at least the human communities that maintain it. Protected areas can both protect species diversity and benefit indigenous communities (Larsen et al. 2012, Dudley et al. 2014). However, they are vulnerable to negative impacts from adjacent nonprotected areas and are difficult to make large enough to encompass the year-round needs of many species (Janzen 1986, Tjørve 2010). Small islands are often suitable for whole-island protected areas (Spatz et al. 2014), which greatly reduces the impact of most external threats not associated with climate change. Furthermore, most island endemic species have evolved without large terrestrial migrations and can therefore persist in smaller protected areas as long as they encompass entire islands or significant portions of an island.

Reintroductions of locally extinct species can restore lost ecosystem functions and ecosystem services that benefit local communities (Zavaleta et al. 2009, Sangha et al. 2011). It is particularly important on islands because the simplified ecosystems make the role of each individual species more

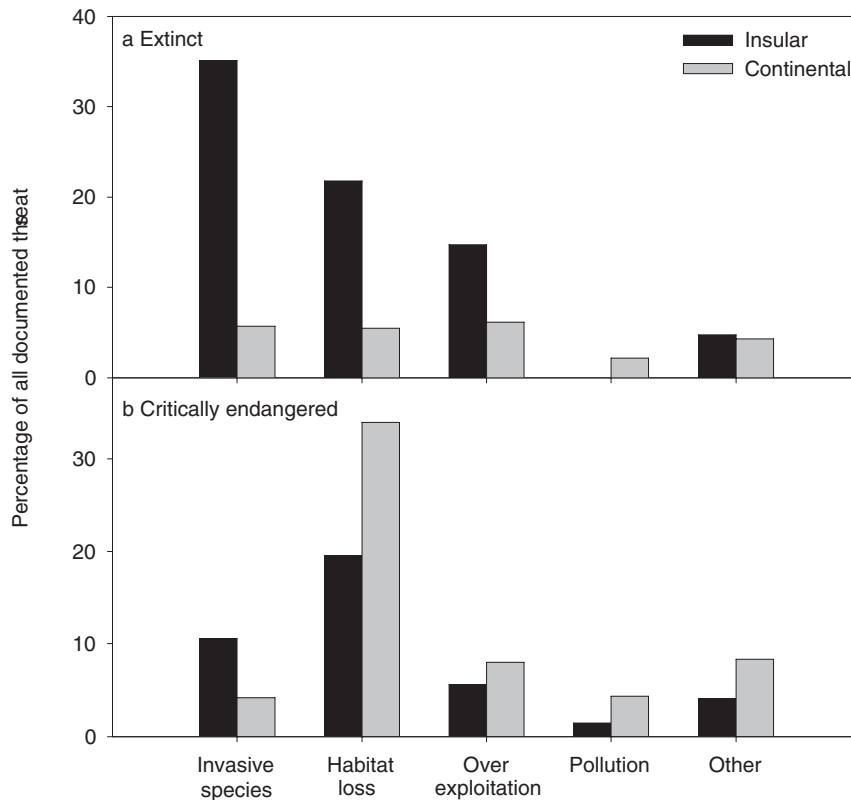


Figure 3. The relative importance of different threats for the decline of all extinct species (a) and all critically endangered species (b) confined to islands (insular) and species occurring primarily on continents (continental) from all taxa in IUCN (2010). The other category includes climate change, disease, and other threats.

significant (Aslan et al. 2013). Furthermore, reintroductions are perhaps more likely to be successful in the simplified ecosystems of small islands because the empty niche of a locally extinct species is less likely to be filled by other species. For these same reasons, when the original species has gone extinct, taxon substitutions are also both more necessary and more feasible on islands (Atkinson 2001).

Eradicating damaging invasive species is an effective and increasingly widespread native species conservation tool for which techniques are improving steadily and are ready to be applied to larger islands with more dense human populations (Howald et al. 2007, Keitt et al. 2011). Invasive rodents (e.g., *Rattus rattus* and *Rattus norvegicus*) are estimated to occur on 80% of islands worldwide and to have played a role in 40% of all bird extinctions (Atkinson 1985). Invasive rodents also compete directly with rural people for food; damage property; spread diseases, such as leptospirosis; and suppress native species that provide valuable ecosystem services (Mwebaze et al. 2010, Banks and Hughes 2012). Feral cats (*Felis silvestris catus*) are common on human-inhabited islands (Fitzgerald 1988) and have played a role in 14% global bird, mammal, and reptile extinctions (Medina et al. 2011). Feral cats also represent a human disease threat via toxoplasmosis (Dabritz and Conrad 2010), leading to increased

probability of schizophrenia (Webster et al. 2006). The eradication of introduced rodents, feral cats, and other invasive species can therefore lead to an ecosystem recovery that can benefit both biodiversity and island peoples (Aguirre-Muñoz et al. 2008, Keitt et al. 2011). Improved livelihoods can facilitate a community's ability to maintain or revitalize their culture, including indigenous languages. Globally, thousands of islands, many of which are home to indigenous cultures, could benefit from invasive species eradications. For example, the recent invasion of New Guinea Island by macaques (*Macaca fascicularis*) threatens the livelihoods of indigenous communities and the survival of native species. This is significant because New Guinea houses an estimated 15% of all languages and 5% of all terrestrial species (Kemp and Burnett 2003, Lewis et al. 2013).

The results of our review demonstrate the disproportionate importance of islands to biological and linguistic diversity. Furthermore, they suggest that proven approaches to biodiversity conservation are particularly effective when applied to islands and can have the potential to benefit indigenous human populations and their languages.

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References cited

- Aguirre-Muñoz A, et al. 2008. High-impact conservation: Invasive mammal eradications from the islands of Western Mexico. *AMBIO* 37: 101–107.
- Amori G, Gippoliti S, Helgen KM. 2008. Diversity, distribution, and conservation of endemic island rodents. *Quaternary International* 182: 6–15.
- Arnold M, Powell B, Shanley P, Sunderland, TCH. 2011. Editorial: Forests, biodiversity and food security. *International Forestry Review* 13: 259–264.
- Aslan CE, Zavaleta ES, Tershy B, Croll D. 2013. Mutualism disruption threatens global plant biodiversity: A systematic review. *PLOS ONE* 8 (art. e66993).
- Atkinson IAE. 1985. The spread of commensal species of *Rattus* to oceanic islands and their effects on island avifaunas. Pages 35–81 in Moors PJ, ed. *Conservation of Island Birds: Case Studies for the Management of Threatened Island Species*. International Council for Bird Preservation. Technical Publication no. 3.
- Atkinson IAE. 2001. Introduced mammals and models for restoration. *Biological Conservation* 99: 81–96.
- Austin PK, Sallabank J, eds. 2011. *The Cambridge Handbook of Endangered Languages*. Cambridge University Press.
- Banks PB, Hughes NK. 2012. A review of the evidence for potential impacts of black rats (*Rattus rattus*) on wildlife and humans in Australia. *Wildlife Research* 39: 78–88.

- Blumstein DT, Daniel JC. 2005. The loss of anti-predator behaviour following isolation on islands. *Proceedings of the Royal Society B* 272: 1663–1668.
- Bowen L, Van Vuren D. 1997. Insular endemic plants lack defenses against herbivores. *Conservation Biology* 11: 1249–1254.
- Brooks TM, Mittermeier RA, da Fonseca GA, Gerlach J, Hoffmann M, Lamoureux JF, Mittermeier CG, Pilgrim JD, Rodrigues ASL. 2006. Global biodiversity conservation priorities. *Science* 313: 58–61.
- Dabritz HA, Conrad PA. 2010. Cats and *Toxoplasma*: Implications for public health. *Zoonoses Public Health* 57: 34–52.
- Dudley N, MacKinnon K, Stolton S. 2014. The role of protected areas in supplying ten critical ecosystem services in drylands: A review. *Biodiversity* 15: 178–184.
- Fitzgerald BM. 1988. Diet of domestic cats and their impact on prey populations. In Turner DC, Bateson P, eds. *The Domestic Cat: The Biology of Its Behaviour*. Cambridge University Press.
- Frankham R. 1997. Do island populations have less genetic variation than mainland populations? *Heredity* 78: 311–327.
- Gorenflo LJ, Romaine S, Mittermeier RA, Walker-Painemilla K. 2012. Co-occurrence of linguistic and biological diversity in biodiversity hotspots and high biodiversity wilderness areas. *Proceedings of the National Academy of Sciences* 109: 8032–8037.
- Halliday MAK. 2001. New ways of meaning: The challenge to applied linguistics. Pages 175–202 in Fill A, Mühlhäusler P, eds. *The Ecological Reader: Language, Ecology, and Environment*. Continuum.
- Hedges SB, Conn CE. 2012. A new skink fauna from Caribbean islands (Squamata, Mabuyidae, Mabuyinae). *Zootaxa* 3288: 1–244.
- Howald G, et al. 2007. Invasive rodent eradication on islands. *Conservation Biology* 21: 1258–1268.
- [IUCN] International Union for Conservation of Nature. 2010. The IUCN Red List of Threatened Species. IUCN. (2 April 2010; www.iucnredlist.org)
- Janzen DH. 1986. The eternal internal threat. Pages 286–303 in Soule ME, ed. *Conservation Biology: The Science of Scarcity and Diversity*. Sinauer Associates.
- Kalaba FK, Quinn CH, Dougill AJ. 2013. Contribution of forest provisioning ecosystem services to rural livelihoods in the Miombo woodlands of Zambia. *Population and Environment* 35: 159–182.
- Keitt B, Campbell K, Saunders A, et al. 2011. The Global Islands Invasive Vertebrate Eradication Database: A tool to improve and facilitate restoration of island ecosystems. Pages 74–77 in Veitch CR, Clout MN, Towns DR, eds. *Island Invasives: Eradication and Management*. International Union for Conservation of Nature.
- Kemp NJ, Burnett JB. 2003. Final Report: A Biodiversity Risk Assessment and Recommendations for Risk Management of Long-Tailed Macaques (*Macaca fascicularis*) in New Guinea. Indo-Pacific Conservation Alliance.
- Kier G, et al. 2009. A global assessment of endemism and species richness across island and mainland regions. *Proceedings of the National Academy of Sciences* 106: 9322–9327.
- Köhler M, Moyà-Solà S. 2009. Physiological and life history strategies of a fossil large mammal in a resource-limited environment. *Proceedings of the National Academy of Sciences* 106: 20354–20358.
- Larson FW, Turner WR, Brooks TM. 2012. Conserving critical sites for biodiversity provides disproportionate benefits to people. *PLOS ONE* 7 (art. e36971).
- Lewis MP, ed. 2009. *Ethnologue: Languages of the World*. 16th ed. SIL International.
- Lewis MP, Simons GF, Fennig CD, eds. 2013. *Ethnologue: Languages of the World*, 17th ed. SIL International.
- Loehle C, Eschenbach W. 2012. Historical bird and terrestrial mammal extinction rates and causes. *Diversity Distribution* 18: 84–91.
- MacArthur RH, Wilson EO. 1967. *The Theory of Island Biogeography*. Princeton University Press.
- Maffi L. 2005. Linguistic, cultural, and biological diversity. *Annual Review of Anthropology* 34: 599–617.
- Maffi L. 2007. Biocultural diversity and sustainability. Pages 267–277 in Pretty J, Ball A, Benton T, Guivant J, Lee D, Orr D, Pfeffer M, Ward H, eds. *Sage Handbook on Environment and Society*. Sage.
- Martin TG, et al. 2012. Acting fast helps avoid extinction. *Conservation Letters* 5: 274–280.
- Medina FM, et al. 2011. A global review of the impacts of invasive cats on island endangered vertebrates. *Global Change Biology* 17: 3503–3510.
- Mwebaze P, MacLeod A, Tomlinson D, Barois H, Rijpma J. 2010. Economic valuation of the influence of invasive alien species on the economy of the Seychelles islands. *Ecological Economics* 69: 2614–2623.
- Myers N, Mittermeier RA, Mittermeier CG, Da Fonseca GA, Kent J. 2000. Biodiversity hotspots for conservation priorities. *Nature* 403: 853–858.
- Newton I. 2003. *The Speciation and Biogeography of Birds*. Academic Press.
- Pimm S, Raven P, Peterson A, Şekercioğlu ÇH, Ehrlich PR. 2006. Human impacts on the rates of recent, present, and future bird extinctions. *Proceedings of the National Academy of Sciences* 103: 10941–10946.
- Pimm SL, et al. 2014. The biodiversity of species and their rates of extinction, distribution, and protection. *Science* 344 (art. 1246752).
- Pretty J, et al. 2009. The intersections of biological diversity and cultural diversity: Towards integration. *Conservation Society* 7: 100–112.
- Ramstad KM, et al. 2007. Species and cultural conservation in New Zealand: Maori traditional ecological knowledge of tuatara. *Conservation Biology* 21: 455–464.
- Ricketts TH, et al. 2005. Pinpointing and preventing imminent extinctions. *Proceedings of the National Academy of Sciences* 102: 18497–18501.
- Sangha KK, Butler JRA, Delisle A, Stanley O. 2011. Identifying links between ecosystem services and Aboriginal well-being and livelihoods in north Australia: Applying the Millennium Ecosystem Assessment framework. *Journal of Environmental Science and Engineering* 5: 931–946.
- Spatz DR, et al. 2014. The biogeography of globally threatened seabirds and island conservation opportunities. *Conservation Biology* 28: 1282–1290.
- Sutherland WJ. 2003. Parallel extinction risk and global distribution of languages and species. *Nature* 23: 276–279.
- Terborgh J, et al. 2001. Ecological meltdown in predator-free forest fragments. *Science* 294: 1923–1926.
- Tjørve E. 2010. How to resolve the SLOSS debate: Lessons from species-diversity models. *Journal of Theoretical Biology* 264: 604–612.
- [UNEP-WCMC] United Nations Environment Programme World Conservation Monitoring Centre. 2013. Global Islands Database. UNEP-WCMC. (30 January 2013; www.unep-wcmc.org)
- Webster J, Lamberton P, Donnelly C, Torrey E. 2006. Parasites as causative agents of human affective disorders? The impact of anti-psychotic, mood-stabilizer and anti-parasite medication on *Toxoplasma gondii*'s ability to alter host behaviour. *Proceedings of the Royal Society B* 273: 1023–1030.
- Weigelt P, Jetz W, Kreft H. 2013. Bioclimatic and physical characterization of the world's islands. *Proceedings of the National Academy of Sciences* 110: 15307–15312.
- Whittaker RJ, Fernández-Palacios JM. 2007. *Island Biogeography: Ecology, Evolution, and Conservation*. Oxford University Press.
- Wurm SA. 2003. The language situation and language endangerment in the Greater Pacific area. Pages 15–48 in Janse M, Tol S, eds. *Language Death and Language Maintenance: Theoretical, Practical and Descriptive Approaches*. Current Issues in Linguistic Theory, vol. 240. John Benjamins.
- Zavaleta E, et al. 2009. Ecosystem responses to community disassembly. *Annals of the New York Academy of Sciences* 1162: 311–333.

Bernie R. Tershy (tershy@ucsc.edu) is an adjunct professor, Kuo-Wei Shen is a graduate student, Kelly M. Newton is a research specialist, and Donald A. Croll is a professor with the Coastal Conservation Action Lab, in the Department of Ecology and Evolutionary Biology at the University of California, Santa Cruz. Nick Holmes is the director of science at Island Conservation, in Santa Cruz, California.